

MOVEMENT ECOLOGY

Interacting effects of human presence and landscape modification on birds and mammals

Ruth Y. Oliver^{1,2,3,*†}, Scott W. Yanco^{1,2,4,5,*†}, Diego Ellis-Soto^{1,2,†}, Brett R. Jesmer⁶, Juliet Cohen³, Song Gao⁷, Robert Patchett⁸, Tal Avgar^{9,10,11}, Keith Bildstein¹², Nicholas W. Bakner¹³, David Barber¹², Kristin Barker¹⁴, Joseph G. Barnes¹⁵, Guillaume Bastille-Rousseau¹⁶, Jerrold L. Belant¹⁷, John F. Benson¹⁸, Joël Bêty¹⁹, Dean E. Beyer Jr.¹⁷, David Bird^{20,21}, Nathaniel Bowersock²², Andy J. Boyce⁵, Ben S. Carlson^{1,2}, Michael L. Casazza²³, Michael J. Chamberlain²⁴, Michael J. Cherry²⁵, Bret A. Collier²⁶, Alyson Courtemanch²⁷, Sarah C. Davidson^{28,29,30}, Darren DeBloois³¹, Vickie DeNicola³², Christopher R. DeSorbo³³, Robert C. Dowler³⁴, Daniel Dupont³⁵, L. Mark Elbroch³⁶, John Elliott^{21,37,38}, Betsy A. Evans³⁹, W. Mark Ford⁴⁰, David Hancock²¹, Molly Hardesty-Moore⁴¹, Jason E. Hawley⁴², Mackenzie R. Jeffress⁴³, Scott Jennings⁴⁴, Matthew J. Kauffman⁴⁵, Roland Kays^{46,47}, Marcella J. Kelly⁶, Bryan M. Kluever³⁹, Myles Lamont^{21,48}, Scott LaPoint^{49,50}, Tayler N. LaSharr⁵¹, Josee Lefebvre⁵², Pierre Legagneux^{53,54}, Matthias-Claudio Loretto^{28,55}, David Lumpkin⁴⁴, Lindsay A. Martinez⁵⁶, John M. Marzluff⁵⁷, Douglas McCauley⁵⁸, Fiona McDuie^{23,59}, Tony W. Mong⁶⁰, Kevin L. Monteith⁵¹, Thomas Mueller^{61,62}, Levi Newediuk⁶³, Anna C. Ortega⁵⁶, Federico Ossi^{64,65}, Cory Overton²³, J. Clint Perkins⁶⁶, Tyler R. Petroelje⁶⁷, Laura Prugh⁵⁷, Kimberly A. Sager-Fradkin⁶⁸, Michael Seer²¹, Avery L. Shawler¹⁴, Shannon Skalos^{23,69}, Rachel A. Smiley⁵¹, Julia Sommerfeld⁷⁰, Daniel R. Stahler⁷¹, John A. Stephenson⁷², Richard D. Stevens^{66,73}, Nathan J. Svoboda⁷⁴, Jean-Francois Therrien¹², Phillippe J. Thomas⁷⁵, Meredith VanAcker^{50,76,77}, Eric Vander Wal⁶³, Dan E. Varland⁷⁸, Tana L. Verzuh⁷⁹, Brittany L. Wagler⁵¹, Nils Warnock⁴⁴, Stephen L. Webb^{80,81}, Christopher K. Williams⁸², Christopher C. Wilmers⁸³, David W. Wolfson⁸⁴, Julie K. Young⁹, Christian Rutz⁸, Walter Jetz^{1,2}

Sustainable human–wildlife coexistence requires a mechanistic understanding of the many ways that humans affect animals. However, progress is hampered by the lack of accessible data measuring the dynamic presence of people. Here, we leverage mobile-device data to disentangle how human presence and landscape modification differentially influence the use of geographic and environmental space for 37 mammal and bird species across the United States. Human presence affected more than 65% of species, with substantial variation across species. For ~60% of species that responded to human activities, the effects were interdependent—animals tended to react more strongly to human presence in less modified habitats. Our results demonstrate that human presence and landscape modification have complex combined effects on wildlife, which need to be considered for effective management.

Human activities are driving catastrophic loss of biodiversity around the world by altering the climate and reshaping land- and seascapes (1–3). As the march of human encroachment continues, a mechanistic understanding of how humans and wildlife may coexist is critical to maintaining biodiversity (4). Doing so requires charting the many interacting ways in which human activities influence wildlife. Typically, the intensity of landscape modification through urbanization, agriculture, transportation,

and energy infrastructure is assumed to be a robust proxy of human impacts [e.g., (5–8)]. However, a rich theoretical literature predicts that the physical presence of humans themselves can also substantially affect how animals perceive risk, with major consequences for their use of space and resources and interactions with other species (9–11).

The dynamic presence of humans and their vehicles is rarely considered in ecological studies, presumably because data at sufficiently fine spatial and temporal scales are largely inaccessible (12). Therefore, whether human presence is a major driver of wildlife behavior has only been directly tested in local settings, where data on humans and wildlife can sometimes be collected in tandem, or at regional scales using coarse proxies of human presence, such as pandemic-related lockdowns [e.g., (13–15)]. Moreover, recent research has primarily focused on mammals, with little work investigating the effects of human presence on other taxonomic groups. Finally, whether multiple forms of human activities have additive or interacting influences on wildlife has not been well established (16, 17). Failing to explicitly quantify the presence of humans as an independent driver may obfuscate the ways in which humans affect animal behavior, physiology, demography, and ultimately population viability (18, 19).

We investigated how two major components of human activity—human presence and landscape modification—affect the use of geographic and environmental space by mammals and birds across the continental United States of America (US) (Fig. 1A). We focused our study during the years 2019 and 2020, to leverage high-resolution human-mobility data, which measure daily human presence, that had been temporarily available to researchers during the COVID-19 pandemic (20, 21). The policies to restrict human movement during the early pandemic period also disrupted the typical positive association between human presence and landscape modification, thus decoupling the usually confounded effects of these drivers (22–25).

Animals' use of geographic and environmental space reflects an optimized trade-off between energy expenditure and acquisition and may vary on the basis of species' traits and life-history strategies (26, 27). Human activities may cause animals to use more area as they take less direct routes to avoid contact or explore alternative habitats (28, 29). Alternatively, modifications to landscapes can render them impermeable to wildlife (8, 15, 30), constraining the amount of area available for use, or animals may cover less area as they exploit concentrated resource pockets of anthropogenic food sources (31). Changes in the size of animals' environmental niches indicate a shift in the resources or habitats used, or more generally, a change in exposure to environmental conditions (32, 33). Therefore, human activities may narrow animals' environmental niches by constraining the habitats they can access or broaden their niches by disrupting access to preferred habitats and necessitating the use of marginal ones. Alternatively, human activities could change niche size without a corresponding change in area size by altering the locations individuals used, indicating that species either switched to alternative habitats or were pushed into suboptimal ones, but these alternative explanations cannot be distinguished without information on individuals' fitness. Because human presence and landscape modification can affect wildlife through multiple mechanisms that depend on species' real and perceived risk, we predicted that responses to each component would vary strongly by species and environmental context.

We estimated the daily presence of humans, normalized by area, using mobile-device counts (SafeGraph; [safegraph.com](https://www.safegraph.com)) available at the US Census block group level and landscape modification based on a multiyear aggregated metric of anthropogenic modification within 1 km² pixels (Fig. 1A and fig. S5) (34). We used locations recorded by animal-mounted GPS devices to derive two metrics for each individual on a weekly basis from January through August, inclusive, of both years (Fig. 1A and table S1). First, we used dynamic Brownian bridge movement models to estimate the amount of geographic space used by individual animals on a weekly basis (hereafter “area size”; figs. S1

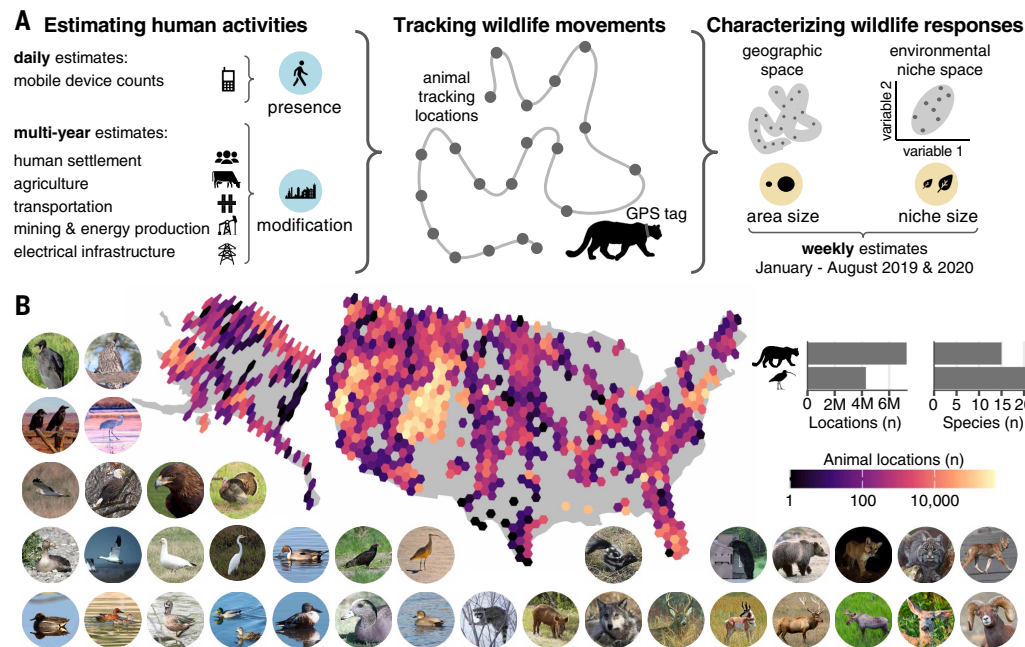


Fig. 1. Wildlife responses to the major components of human activity across the United States. (A) The impact of human activities was quantified by linking daily estimates of human presence and multiyear estimates of landscape modification, respectively, to locations of GPS-tagged animals. GPS locations were used to estimate animals' weekly area and niche size (36). (B) Responses were examined for 37 bird and mammal species across the continental United States, based on ~11.8 million animal locations from 4581 individuals. [Image credits: Plains spotted skunk, Clint Perkins; Great egret, Nils Warnock; all other images, Creative Commons Zero (CCO) or public domain (supplementary text)]

¹Department of Ecology and Evolutionary Biology, Yale University, New Haven, CT, USA. ²Center for Biodiversity and Global Change, Yale University, New Haven, CT, USA. ³Bren School of Environmental Science and Management, University of California Santa Barbara, Santa Barbara, CA, USA. ⁴School of Environment and Sustainability, University of Michigan, Ann Arbor, MI, USA. ⁵Migratory Bird Center, Smithsonian's National Zoo and Conservation Biology Institute, Washington, DC, USA. ⁶Department of Fish and Wildlife Conservation, Virginia Tech, Blacksburg, VA, USA. ⁷Department of Geography, University of Wisconsin-Madison, Madison, WI, USA. ⁸Centre for Biological Diversity, School of Biology, University of St Andrews, St Andrews, UK. ⁹Department of Wildland Resources and the Ecology Center, Utah State University, Logan, UT, USA. ¹⁰Department of Biology, University of British Columbia, Okanagan, BC, Canada. ¹¹Biodiversity Pathways Ltd., Calgary, AB, Canada. ¹²Hawk Mountain Sanctuary, Orwigsburg, PA, USA. ¹³Waterfowl and Upland Gamebird Center, University of Delaware, Newark, DE, USA. ¹⁴Department of Environmental Science, Policy, and Management, University of California Berkeley, Berkeley, CA, USA. ¹⁵US Fish and Wildlife Service, Migratory Bird Program, Reno, NV, USA. ¹⁶Cooperative Wildlife Research Lab, Southern Illinois University, Carbondale, IL, USA. ¹⁷Department of Fisheries and Wildlife, Michigan State University, East Lansing, MI, USA. ¹⁸School of Natural Resources, University of Nebraska-Lincoln, Lincoln, NE, USA. ¹⁹Département de Biologie et Centre d'études nordiques, Université du Québec à Rimouski, Rimouski, QC, Canada. ²⁰Department of Natural Resource Sciences, McGill University, West Montreal, QC, Canada. ²¹Hancock Wildlife Foundation, Surrey, BC, Canada. ²²Missouri Department of Conservation, Columbia, MO, USA. ²³United States Geological Survey, Western Ecological Research Center, Dixon Field Station, Dixon, CA, USA. ²⁴Warnell School of Forestry and Natural Resources, University of Georgia, Athens, GA, USA. ²⁵Cesar Kleberg Wildlife Research Institute, Texas A&M University-Kingsville, Kingsville, TX, USA. ²⁶School of Renewable Natural Resources, Louisiana State University, Baton Rouge, LA, USA. ²⁷Wyoming Game and Fish Department, Jackson, WY, USA. ²⁸Max Planck Institute of Animal Behavior, Department of Animal Migration, Radolfzell, Germany. ²⁹The Ohio State University, Department of Civil, Environmental and Geodetic Engineering, Columbus, USA. ³⁰Department of Biology, University of Konstanz, Konstanz, Germany. ³¹Division of Wildlife Resources, Utah Department of Natural Resources, Salt Lake City, UT, USA. ³²Field Engine Wildlife Research and Management, East Haddam, CT, USA. ³³Biodiversity Research Institute, Portland, ME, USA. ³⁴Department of Biology, Angelo State University, San Angelo, TX, USA. ³⁵Département des sciences expérimentales, Université de Saint-Boniface, Winnipeg, MB, Canada. ³⁶Panthera, New York, NY, USA. ³⁷Environment and Climate Change Canada, Delta, BC, Canada. ³⁸Applied Animal Biology, University of British Columbia, Vancouver, BC, Canada. ³⁹US Dept of Agriculture, Animal and Plant Health Inspection Service, Wildlife Services, National Wildlife Research Center, Gainesville, FL, USA. ⁴⁰United States Geological Survey Virginia Cooperative Fish and Wildlife Research Unit, Department of Fish and Wildlife Conservation, Virginia Tech, Blacksburg, VA, USA. ⁴¹Sierra College, Rocklin, CA, USA. ⁴²Connecticut Department of Environmental Protection, North Franklin, CT, USA. ⁴³Nevada Department of Wildlife, Elko, NV, USA. ⁴⁴All Hands Ecology, Marshall, CA, USA. ⁴⁵US Geological Survey, Wyoming Cooperative Fish and Wildlife Research Unit, Department of Zoology and Physiology, University of Wyoming, Laramie, WY, USA. ⁴⁶North Carolina Museum of Natural Sciences, Raleigh, NC, USA. ⁴⁷Department of Forestry and Environmental Resources, North Carolina State University, Raleigh, NC, USA. ⁴⁸TerraFauna Wildlife Consulting, Inc, Surrey, BC, Canada. ⁴⁹Black Rock Forest, Cornwall, NY, USA. ⁵⁰Department of Ecology, Evolution, and Environmental Biology, Columbia University, New York, NY, USA. ⁵¹Haub School of Environment and Natural Resources, Wyoming Cooperative Fish and Wildlife Research Unit, Department of Zoology and Physiology, University of Wyoming, Laramie, WY, USA. ⁵²Canadian Wildlife Service, Environment and Climate Change Canada, Québec Region, QC, Canada. ⁵³Département de biologie & Centre d'Études Nordiques, Université Laval, Québec, QC, Canada. ⁵⁴Centre d'Études Biologiques de Chizé, Université de La Rochelle, Villiers en Bois, France. ⁵⁵Research Institute of Wildlife Ecology, Department of Interdisciplinary Life Sciences, University of Veterinary Medicine Vienna, Vienna, Austria. ⁵⁶Wyoming Cooperative Fish and Wildlife Research Unit, Department of Zoology and Physiology, University of Wyoming, Laramie, WY, USA. ⁵⁷School of Environmental and Forest Sciences, University of Washington, Seattle, WA, USA. ⁵⁸Department of Ecology, Evolution, and Marine Biology, University of California Santa Barbara, Santa Barbara, CA, USA. ⁵⁹San Jose State University Research Foundation, San Jose State University, San Jose, CA, USA. ⁶⁰Wyoming Game and Fish Department, Cheyenne, WY, USA. ⁶¹Senckenberg Biodiversity and Climate Research Centre (SBiK-F), Frankfurt am Main, Germany. ⁶²Department of Biological Sciences, Goethe University Frankfurt, Frankfurt am Main, Germany. ⁶³Department of Biology, Memorial University of Newfoundland, St. John's, NL, Canada. ⁶⁴Animal Ecology Unit, Research and Innovation Centre, Fondazione Edmund Mach, Trento, Italy. ⁶⁵National Biodiversity Future Center (NBFC), Palermo, Italy. ⁶⁶Department of Natural Resources Management, Texas Tech University, Lubbock, TX, USA. ⁶⁷Michigan Department of Natural Resources, Marquette, MI, USA. ⁶⁸Lower Elwha Klallam Tribe, Port Angeles, WA, USA. ⁶⁹United States Geological Survey, Eastern Ecological Science Center, Patuxent Research Refuge, Laurel, MD, USA. ⁷⁰Büro für faunistische Fachfragen, Linden, Germany. ⁷¹Yellowstone National Park, National Park Service, WY, USA. ⁷²Grand Teton National Park, National Park Service, WY, USA. ⁷³Natural Science Research Laboratory of the Museum at Texas Tech University, Lubbock, TX, USA. ⁷⁴Alaska Department of Fish and Game, Kodiak, AK, USA. ⁷⁵Environment and Climate Change Canada, National Wildlife Research Centre, Carleton University, Ottawa, ON, Canada. ⁷⁶Global Health Program, Smithsonian's National Zoo and Conservation Biology Institute, Washington, DC, USA. ⁷⁷Department of Evolution, Ecology and Organismal Biology, University of California, Riverside, Riverside, CA, USA. ⁷⁸Coastal Raptors, Hoquiam, WA, USA. ⁷⁹School of Computing, Haub School of Environment and Natural Resources, University of Wyoming, Laramie, WY, USA. ⁸⁰Natural Resources Institute, Texas A&M University, College Station, TX, USA. ⁸¹Department of Rangeland, Wildlife, and Fisheries Management, Texas A&M University, College Station, TX, USA. ⁸²Department of Entomology and Wildlife Ecology, University of Delaware, Newark, DE, USA. ⁸³Environmental Studies Department, University of California, Santa Cruz, CA, USA. ⁸⁴Fisheries, Wildlife, and Conservation Biology, University of Minnesota, St. Paul, MN, USA. ***Corresponding author. rutholiver@ucsb.edu (R.Y.O.); yancos@si.edu (S.W.Y.)** †These authors contributed equally to this work. ‡Present address: Department of Environmental Science Policy and Management, University of California, Berkeley, CA, USA §Present address: Department of Wildland Resources, S.J. and Jessie E. Quinney College of Agriculture and Natural Resources, Utah State University, Logan, UT, USA. ¶Present address: Department of Biology, Mount Royal University, Calgary, AB, Canada.

and S6 and table S3) (35, 36). Second, we estimated the weekly environmental niche size of individual animals, defined as the pooled variance of multidimensional hypervolumes of the set of environmental conditions across recorded GPS locations (hereafter “niche size”; fig. S2) (36, 37). We estimated responses of area and niche size to human presence and landscape modification for birds ($n = 22$ species) and mammals ($n = 15$ species) based on ~11.8 million animal-tracking locations ($n = 4581$ individuals) using species-specific Bayesian generalized linear mixed effects models. All models controlled for environmental conditions (i.e., vegetation productivity and air temperature) and for area size in niche size models (Fig. 1B and table S2) (36).

Interacting effects of human activities

Human presence was associated with changes in area or niche size for 67% of the mammal species and 68% of the bird species studied (Fig. 2A). More than half of species (57%) were affected by both human presence and landscape modification (Fig. 2B; area size: 46%; niche size: 27%). For most species (59%) that responded to human activity, the effects of human presence and landscape modification on area or

niche size were interactive (i.e., the response to one driver depended on the strength of the other) (Fig. 2A). Altogether, our findings show that species responded to human presence and landscape modification by changing the amount of area used, adjusting the size of their environmental niches, or both (Fig. 2C). These results demonstrate that directly quantifying the presence of humans over short timescales is necessary to fully capture the impact of human activity on wildlife.

Most mammal species (67%) contracted the amount of area used in response to human presence or landscape modification (Fig. 2A). Many (40%) mammal species that responded to human presence reacted more strongly in less modified habitats, suggesting that some populations inhabiting more developed areas are either habituated to humans or cannot further reduce the amount of area they use while also maintaining access to critical resources (17). Gray wolves (*Canis lupus*) are an exception and expanded the amount of area they used in response to both human presence and landscape modification. Gray wolves have faced considerable persecution throughout North America and are considered to be particularly sensitive to humans (38). Their expanded use of space could reflect a requirement to cover more area to avoid human activities (28, 29).

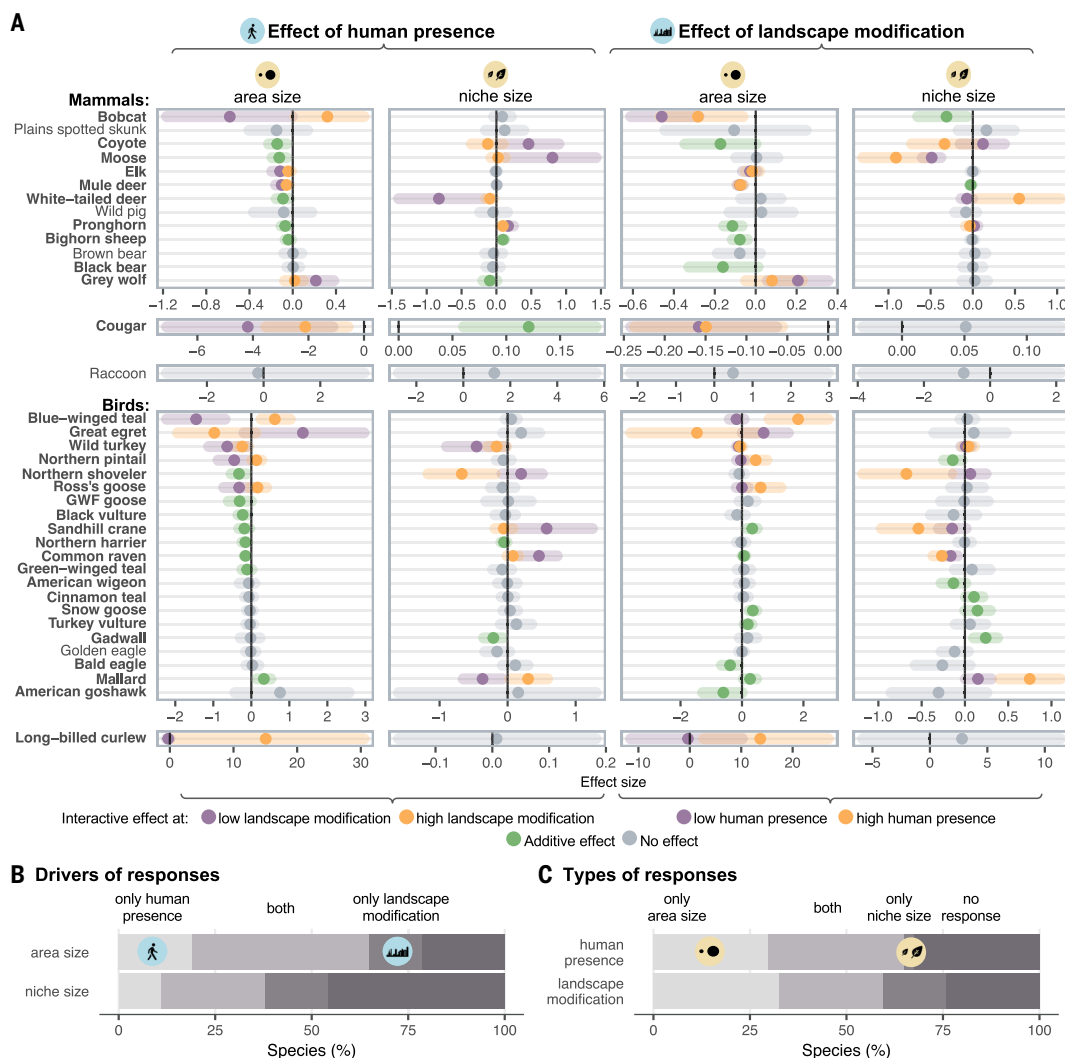


Fig. 2. Interacting effects of human activities on wildlife's use of geographic and environmental space. (A) Effects of human presence and landscape modification on area and niche size used by animals on a weekly timescale. Points show effect-size estimates with 95% credible intervals. Species names in bold indicate species with at least one significant response. Full model results and sample sizes can be found in table S2 and figs. S7 to S80. Effect sizes for cougars, raccoons, and long-billed curlews are shown separately, on different scales. **(B)** The majority of species responded to both human presence and landscape modification. **(C)** Species responded to human presence and landscape modification by changing their area size, niche size, or both.

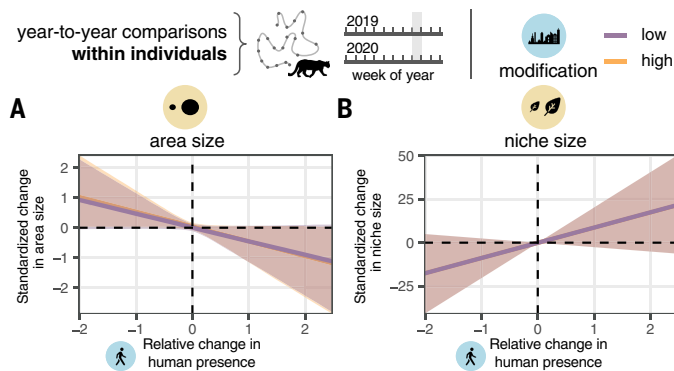


Fig. 3. Plastic behavioral responses to human presence. Conditional effects of changes in human presence on the changes in (A) area and (B) niche size for individual animals on a weekly basis compared between 2019 and 2020. Shading indicates 95% credible intervals. The credible intervals for high- and low-modification estimates are overlapping because of a lack of a significant global interactive effect. Sample size: $n = 19,453$ individual-week pairs, 1251 individuals, 13 species.

The presence of humans also contracted the amount of area used by 41% of the bird species studied (Fig. 2A). For many of these species, the magnitude and direction of the effect of human presence depended on the degree of landscape modification. For example, human activities restricted the amount of area used by wild turkeys (*Meleagris gallopavo*), but they responded less strongly to human presence in more modified environments. By contrast, the direction of the response to human presence for great egrets (*Ardea alba*) depended on the degree of landscape modification. Species that rely on patchily distributed habitats may be less likely to switch patches in highly developed environments, whereas dispersal may be more likely in relatively undeveloped contexts (39).

When controlling for weekly area size in niche size models (36), we found that, for many species, human presence and landscape modification had opposing effects on environmental niches (Fig. 2A). However, the direction of responses varied across species for both birds and mammals. For example, human presence and landscape modification had opposing effects on the niche sizes of white-tailed deer (*Odocoileus virginianus*) and sandhill cranes (*Grus canadensis*), albeit with different directions of responses. White-tailed deer, a highly human-affiliated species in the US, expanded their environmental niches with increasing landscape modification but contracted their niches with increasing human presence (40). Sandhill cranes, a non-human-affiliated species, showed the opposite response. These responses may reflect divergent responses to humans where synanthropic species integrate modified environments into their environmental niches, whereas less habituated species rely on increasingly isolated patches of habitat.

Behavioral plasticity to changes in human presence

To test whether species-level responses were driven by behavioral plasticity of individual animals to year-to-year differences in human activities or differential responses among individuals within species, we used locations from individual animals that were tracked in both 2019 and 2020. Using Bayesian mixed effects models to control for environmental conditions and sampling design, we compared weekly year-to-year differences to estimate the impact of relative changes in human presence on area and niche size of individual animals (36).

As with our species-level analysis, we found that year-to-year differences in human presence were generally negatively associated with changes in area size and positively associated with changes in niche size for individual animals, although species-specific responses varied (Fig. 3 and fig. S3; area size marginal effect at median landscape modification: $\beta = -0.48$, 95% credible interval (CI) = $[-1.37, 0.08]$, Bayesian probability of direction (pd) = 96.0%; niche size marginal effect at

median landscape modification: $\beta = 13.03$, 95% CI = $[8.84, 32.88]$, pd = 90.7%). Because we observed that individual animals responded to changes in human activities, these results provide evidence that the observed species-level responses are at least partially driven by the behavioral plasticity of individual animals as opposed to differential responses among individuals. We found little support for a global interactive effect between changes in human presence and landscape modification in either area size ($\beta = -0.08$, 95% CI = $[-0.41, 0.22]$, pd = 74.5%) or niche size ($\beta = 0.05$, 95% CI = $[-0.45, 0.34]$, pd = 65.4%). However, as with the main effect of human presence, species varied in their responses, with some showing an interactive effect between human presence and landscape modification (fig. S4). As with our species-level analysis, these results suggest that the interacting effects of human presence and landscape modification are not consistent across species.

Combined impact of human activities on wildlife

To estimate the combined impact of human presence and landscape modification on wildlife area and niche sizes, we compared predicted area and niche sizes at low and high levels of human activity for species that showed significant responses to human activity (20% and 80% species-specific quantiles of observed human presence and landscape modification). We restricted predictions to species that showed significant responses because predictions for species without responses would not differ under low and high levels of human activity.

Of species with significant responses to human activities, mammals reduced the amount of area that they used on a weekly basis by a median of 11% per animal (range = $[-35.9, 15.7 \text{ km}^2]$) and birds exhibited a median decrease of 1.8% per animal (range = $[-171.4, 72.6 \text{ km}^2]$) (Fig. 4). However, there was substantial variation among species. For example, we estimated that human activities caused coyotes (*Canis latrans*) to use 11.3 km^2 less area per week per animal, whereas common ravens (*Corvus corax*) used 26 km^2 more area per week per animal. These seemingly alternate responses may be driven by the same mechanism, as both coyotes and ravens are known to exploit anthropogenic food sources (28, 41, 42). However, coyotes face substantial persecution by humans and therefore may do so by restricting their territories, whereas ravens may be emboldened to cover larger areas to access anthropogenic subsidies.

We also found that the impact of human activity contracted species' environmental niches by a median of 2.8%, although there appeared to be substantial interspecific variation (mammals: median = -7.7% , range = $[-58.3\%, 40.15\%]$; birds: median = -2.2% , range = $[-63.5\%, 62.1\%]$). For example, we estimated that human activities caused common ravens to narrow their environmental niche by 46%, whereas cougars (*Puma concolor*) broadened theirs by 10%. These results suggest that ongoing human activities have substantial effects on how animals interact with their environment. However, more detailed information on the form of human activities would be needed to precisely define the impact of these effects.

Concluding remarks

Our work demonstrates that for many species, the effects of landscape modification cannot be understood without taking into account the presence of humans, which can both amplify and dampen its effects. These results indicate that human presence must be explicitly considered to appropriately assess anthropogenic impacts on wildlife. Notably, we were unable to interpret whether observed responses represented a cost or subsidy to animals without more detailed information on the nature of human activities (e.g., hiking versus hunting, presence of anthropogenic food) or the fitness consequences of the observed effects. However, the current access model for human-mobility data necessary to explore these nuances largely prevents their use in conservation research (43). Further, the spatial resolution of many human-mobility datasets does not capture the fine-scale behavioral responses necessary to support human-wildlife coexistence.

The identification of multiple types of responses across species suggests that there are distinct pathways by which wildlife species respond

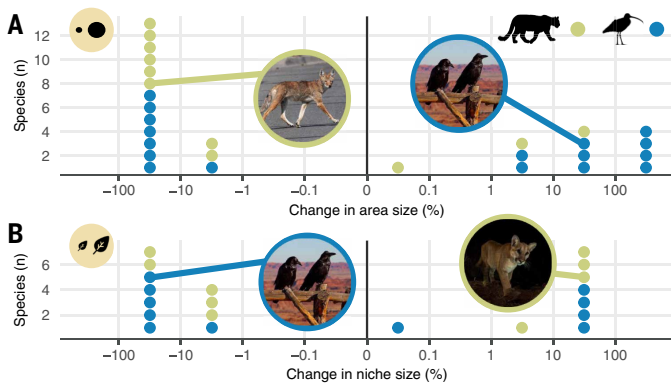


Fig. 4. Combined impact of human activities on wildlife use of geographic and environmental space. Predicted changes in (A) area and (B) niche size based on the difference between high- and low-human activity scenarios for species that showed significant responses to human activity. Points represent the median predicted change in area and niche sizes for an individual animal for each species. Human activity scenarios were estimated on the basis of species-specific 20% (low) and 80% (high) quantiles of observed human presence and landscape modification. Percent change is shown on the logarithmic scale. Species without significant responses are not shown (area size: $n = 22$; niche size: $n = 14$). All images are CC0 or public domain (supplementary text).

to human activities, which could help inform species-specific management. Future work should test the consistency and generality of these pathways across a greater range of taxa as well as how response types relate to population demography (44, 45). The varied responses that we observed across species reflect substantial differences in species' ecology and environmental context, as well as affiliations with humans. Our study was necessarily restricted to species that can be tracked with GPS devices, which excludes most smaller-bodied birds and mammals, which may respond differently to human activities.

Although climate change and habitat loss will continue to place extreme pressure on wildlife populations (46), our results indicate that progress toward human-wildlife coexistence may be achieved through recognizing, and mitigating, the distinct effect of human presence. Conservation benefits could be realized, for example, through the dynamic regulation of human use of wild areas throughout the day and year, as opposed to strict exclusion. For example, it may be possible to manage temporal patterns of vehicular traffic to minimize impacts on wildlife during critical periods or in important environmental refugia (47–49). Global coordination is undoubtedly needed to tackle the biodiversity crisis, but our study has highlighted that there is scope for rapid, local action, using nuanced policies to manage human-wildlife interactions.

REFERENCES AND NOTES

1. P. Jaureguiberry *et al.*, *Sci. Adv.* **8**, eabm9982 (2022).
2. IPBES, Global assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services, Zenodo (2019); <https://doi.org/10.5281/zenodo.6417333>.
3. R. P. Powers, W. Jetz, *Nat. Clim. Chang.* **9**, 323–329 (2019).
4. B. Abrahms *et al.*, *Nat. Clim. Chang.* **13**, 224–234 (2023).
5. N. A. Gilbert, J. L. Stenglein, J. N. Pauli, B. Zuckerberg, *Proc. Natl. Acad. Sci. U.S.A.* **119**, e2206339119 (2022).
6. A. Sempere-Pascual *et al.*, *Nat. Ecol. Evol.* **7**, 1092–1103 (2023).
7. E. O. Aikens, T. B. Wyckoff, H. Sawyer, M. J. Kauffman, *Nat. Ecol. Evol.* **6**, 1733–1741 (2022).
8. S. Mumme *et al.*, *Glob. Chang. Biol.* **29**, 5788–5801 (2023).
9. S. L. Lima, L. M. Dill, *Can. J. Zool.* **68**, 619–640 (1990).
10. J. Berger, *Biol. Lett.* **3**, 620–623 (2007).
11. M. Moleón, J. A. Sánchez-Zapata, *Trends Ecol. Evol.* **38**, 215–216 (2023).
12. D. Ellis-Soto *et al.*, *Nat. Ecol. Evol.* **7**, 1362–1372 (2023).
13. B. A. Nickel, J. P. Suraci, M. L. Allen, C. C. Wilmers, *Biol. Conserv.* **241**, 108383 (2020).
14. M. B. Schrimpf *et al.*, *Sci. Adv.* **7**, eabf5073 (2021).
15. M. A. Tucker *et al.*, *Science* **380**, 1059–1064 (2023).
16. C. C. Wilmers, A. C. Nisi, N. Ranc, *Curr. Biol.* **31**, 3952–3955.e3 (2021).
17. A. C. Burton *et al.*, *Nat. Ecol. Evol.* **8**, 924–935 (2024).
18. Z. Tablado, L. Jenni, *Biol. Rev. Camb. Philos. Soc.* **92**, 216–233 (2017).
19. T. S. Doherty, G. C. Hays, D. A. Driscoll, *Nat. Ecol. Evol.* **5**, 513–519 (2021).
20. S. Chang *et al.*, *Nature* **589**, 82–87 (2021).
21. Y. Kang *et al.*, *Sci. Data* **7**, 390 (2020).
22. C. Rutz *et al.*, *Nat. Ecol. Evol.* **4**, 1156–1159 (2020).
23. C. Rutz, *Nat. Rev. Earth Environ.* **3**, 157–159 (2022).
24. K. M. Gaynor *et al.*, *Front. Ecol. Environ.* **18**, 542–543 (2020).
25. A. E. Bates, R. B. Primack, P. Moraga, C. M. Duarte, *Biol. Conserv.* **248**, 108665 (2020).
26. R. H. MacArthur, E. R. Pianka, *Am. Nat.* **100**, 603–609 (1966).
27. E. L. Charnov, *Theor. Popul. Biol.* **9**, 129–136 (1976).
28. S. D. Gehrt, C. Anchor, L. A. White, *J. Mammal.* **90**, 1045–1057 (2009).
29. S. P. D. Riley *et al.*, *Conserv. Biol.* **17**, 566–576 (2003).
30. M. A. Tucker *et al.*, *Science* **359**, 466–469 (2018).
31. S. D. Newsome, H. M. Garbe, E. C. Wilson, S. D. Gehrt, *Oecologia* **178**, 115–128 (2015).
32. J. Soberón, *Ecol. Lett.* **10**, 1115–1123 (2007).
33. J. Elith, J. R. Leathwick, *Annu. Rev. Ecol. Syst.* **40**, 677–697 (2009).
34. C. M. Kennedy, J. R. Oakleaf, D. M. Theobald, S. Baruch-Mordo, J. Kiesecker, *Glob. Chang. Biol.* **25**, 811–826 (2019).
35. B. Kranstauber, R. Kays, S. D. Lapoint, M. Wikelski, K. Safi, *J. Anim. Ecol.* **81**, 738–746 (2012).
36. See supplementary materials and methods.
37. M. Lu, K. Winner, W. Jetz, *Methods Ecol. Evol.* **12**, 1953–1968 (2021).
38. M. Musiani, P. C. Paquet, *Bioscience* **54**, 50–60 (2004).
39. C. S. Teitelbaum *et al.*, *Mov. Ecol.* **8**, 49 (2020).
40. H. J. Kilpatrick, S. M. Spohr, *Wildl. Soc. Bull.* 1973–2006 **28**, 1023–1029 (2000).
41. C. Ho *et al.*, *Front. Bird Sci.* **2**, 1119507 (2023).
42. P. S. Coates *et al.*, *Biol. Conserv.* **243**, 108409 (2020).
43. R. Y. Oliver *et al.*, *Cell Rep. Sustain.* **1**, 100077 (2024).
44. S. W. Yanco *et al.*, *Trends Ecol. Evol.* **40**, 47–56 (2025).
45. J. P. Suraci *et al.*, *Glob. Chang. Biol.* **27**, 3718–3731 (2021).
46. A. L. Pigot, C. Merow, A. Wilson, C. H. Trisos, *Nat. Ecol. Evol.* **7**, 1060–1071 (2023).
47. C. L. Larson, S. E. Reed, A. M. Merenlender, K. R. Crooks, *PLOS ONE* **11**, e0167259 (2016).
48. J. McGinlay *et al.*, *Forests* **11**, 1214 (2020).
49. J. Whittington, P. Low, B. Hunt, *Sci. Rep.* **9**, 3772 (2019).
50. R. Y. Oliver *et al.*, Supplementary data and code for: Interacting effects of human presence and landscape modification on birds and mammals, Open Science Framework (2025); <https://doi.org/https://doi.org/10.17605/OSF.IO/3UA2C>.
51. D. DeBloois, J. K. Young, T. Avgar, Data from: Study “GPS tracking of cougars in Utah by UDWR (2019–2020),” Movebank (2026); <https://doi.org/10.5441/001/1.712>.
52. T. Avgar, D. DeBloois, D. Mangus, R. Robinson, J. K. Young, Data from: Study “GPS tracking of ungulates in Utah by UDWR (2019–2020),” Movebank (2026); <https://doi.org/10.5441/001/1.711>.
53. G. Bastille-Rousseau, M. Egan, Data from: Study “Illinois white-tailed deer and predators,” Movebank (2026); <https://doi.org/10.5441/001/1.649>.
54. K. L. Bildstein, D. Barber, M. J. Bechard, M. Graña Grilli, J.-F. Therrien, Data from: Study “Vultures Acopian Center USA GPS” (2003–2021), Movebank (2021); <https://doi.org/10.5441/001/1.f3qt46r2>.
55. A. J. Boyce, J. F. Kelly, P. M. Cimprich, Data from: Study “Full annual cycle movements of Long-billed Curlews breeding in Montana USA (Boyce SBCI),” Movebank (2026); <https://doi.org/10.5441/001/1.698>.
56. M. J. Chamberlain, N. W. Bakner, B. A. Collier, Data from: Study “GPS tracking of wild turkeys in the southeastern US,” Movebank (2026); <https://doi.org/10.5441/001/1.671>.
57. M. J. Cherry, M. Kelly, W. M. Ford, Data from: Study “GPS tracking of white-tailed deer in western Virginia,” Movebank (2026); <https://doi.org/10.5441/001/1.718>.
58. M. J. Cherry, M. Kelly, W. M. Ford, Data from: Study “GPS tracking of wild pigs in southern Florida (2019–2020),” Movebank (2026); <https://doi.org/10.5441/001/1.719>.
59. A. B. Courtemanch, Data from: Study “Jackson Moose,” Movebank (2026); <https://doi.org/10.5441/001/1.701>.
60. C. R. DeSorbo, Data from: Study “Bald Eagle - *Haliaeetus leucocephalus* - Biodiversity Research,” Movebank (2026); <https://doi.org/10.5441/001/1.704>.
61. L. M. Elbroch, K. Sager-Fradkin, Data from: Study “Olympic Cougar Project,” Movebank (2026); <https://doi.org/10.5441/001/1.716>.
62. J. G. Barnes, Data from: Study “Golden Eagle Nevada,” Movebank (2026); <https://doi.org/10.5441/001/1.693>.
63. J. G. Barnes, M. R. Jeffress, Data from: Study “GPS tracking of American Goshawks—Nevada Department of Wildlife,” Movebank (2026); <https://doi.org/10.5441/001/1.665>.
64. M. Hardesty-Moore, D. J. McCauley, Data from: Study “UCSB Urban Raccoon Project,” Movebank (2026); <https://doi.org/10.5441/001/1.709>.
65. J. E. Hawley, Data from: Study “American Black bear movement ecology in Connecticut,” Movebank (2026); <https://doi.org/10.5441/001/1.706>.
66. S. Jennings, D. Lumpkin, N. Warnock, E. Condeso, Data from: Study “Audubon Canyon Ranch egret telemetry project,” Movebank (2025); <https://doi.org/10.5441/001/1.641>.
67. R. Kays, M. Engelmann, T. Simons, Data from: Study “LifeTrack Bald Eagle,” Movebank (2026); <https://doi.org/10.5441/001/1.642>.

68. B. M. Kluever, Data from: Study “Vultures tagged in the southeastern United States,” Movebank (2026); <https://doi.org/10.5441/001/1.666>.

69. M. Lamont *et al.*, Data from: Study “Bald Eagle (*Haliaeetus leucocephalus*) in the Pacific Northwest,” Movebank (2026); <https://doi.org/10.5441/001/1.702>.

70. S. LaPoint, Data from: Study “Carnivore movements near Black Rock Forest New York,” Movebank (2026); <https://doi.org/10.5441/001/1.652>.

71. P. Legagneux, J. Lefebvre, C. K. Williams, P. Thomas, J. B  ty, Data from: Study “Greater snow goose migration,” Movebank (2026); <https://doi.org/10.5441/001/1.707>.

72. M.-C. Loretto, M. Wikelski, K. Safi, T. Mueller, J. M. Marzluff, Data from: Study “Common Ravens - Yellowstone,” Movebank (2026); <https://doi.org/10.5441/001/1.653>.

73. L. A. Martinez *et al.*, Data from: Study “Wyoming Bighorn Mountain Moose,” Movebank (2026); <https://doi.org/10.5441/001/1.714>.

74. K. L. Monteith, B. Wagler, R. Smiley, T. LaSharr, Data from: Study “Bighorn sheep and mule deer in Wyoming (2019-2020),” Movebank (2026); <https://doi.org/10.5441/001/1.703>.

75. V. Z  rate *et al.*, Data from: Study “Snowy Range Moose in Wyoming (2019-2020),” Movebank (2026); <https://doi.org/10.5441/001/1.700>.

76. A. Ortega, L. Wilde, K. L. Monteith, M. J. Kauffman, Data from: Study “Wyoming Sublette Mule Deer (2018-2020),” Movebank (2026); <https://doi.org/10.5441/001/1.708>.

77. C. T. Overton, M. L. Casazza, S. M. Skalos, GPS-Marked Waterfowl and Harrier Locations from January 2019 through August 2020 in Western North America, U.S. Geological Survey ScienceBase (2024); <https://doi.org/10.5066/P13E7PCP>.

78. J. C. Perkins, R. D. Stevens, R. C. Dowler, Data from: Study “Spatial ecology of plains spotted skunks in a Texas coastal prairie rangeland (Perkins 2019-2021),” Movebank (2026); <https://doi.org/10.5441/001/1.713>.

79. L. R. Prugh, Data from: Study “GPS tracking of bobcats and coyotes in northern Washington,” Movebank (2023); <https://doi.org/10.5441/001/1.gm93267b>.

80. D. R. Stahler, Data from: Study “Wolves, Stahler, Yellowstone National Park,” Movebank (2026); <https://doi.org/10.5441/001/1.717>.

81. M. VanAcker *et al.*, Resource selection by New York City deer reveals the effective interface between wildlife, zoonotic hazards, and humans, Dryad (2023); <https://doi.org/10.5061/dryad.t76hdr85p>.

82. S. L. Webb, Data from: Study “Wild Pig/Boar GPS Study on Rangeland-Sus scrofa-Oklahoma,” Movebank (2026); <https://doi.org/10.5441/001/1.672>.

83. D. W. Wolfson, D. E. Andersen, J. R. Fieberg, Data from: Study “Sandhill Cranes; Minnesota, USA,” Movebank (2026); <https://doi.org/10.5441/001/1.650>.

84. C. M. Kennedy, J. Oakleaf, D. M. Theobald, S. Baruch-Mordo, J. Kiesecker, Global Human Modification, figshare (2018); <https://doi.org/10.6084/m9.figshare.7283087.v1>.

ACKNOWLEDGMENTS

The authors thank the many individuals that contributed to the collection of the animal tracking datasets included in this study, including D. Andersen, E. Condeso, J. Fieberg, J. Kelly, D. Mangus, and R. Robinson. The authors thank three anonymous reviewers for thoughtful feedback. The opinions, findings, and conclusions or recommendations expressed in this material are those of the authors and do not necessarily reflect the views of the funders. Any use of trade, firm, or product names is for descriptive purposes only and does not imply endorsement by the US Government. **Funding:** The study originated at the Max Planck–Yale Center for Biodiversity Movement and Global Change of the Yale Center for Biodiversity and Global Change, which supported the contributions of R.Y.O., S.W.Y., D.E.S., B.R.J., and W.J. The article is a contribution of the COVID-19 Bio-Logging Initiative, which is funded in part by the Gordon and Betty Moore Foundation (GBMF9881) and the National Geographic Society (NGS-82515R-20) (both grants to C.R.) and endorsed by the United Nations Decade of Ocean Science for Sustainable Development. R.Y.O. acknowledges support from the Kuni Endowed Junior Faculty Fellowship. D.E.S. acknowledges support from NASA FINESST (80NSSC22K1535) and the Yale Institute for Biospheric Studies. S.G. acknowledges support from the National Science Foundation (2027375 and 2112606). S.C.D. acknowledges support from NASA (80NSSC21K1182). Use was made of computational facilities purchased with funds from the National Science Foundation (CNS-1725797) and administered by the Center for Scientific Computing (CSC). The CSC is supported by the California NanoSystems Institute and the Materials Research Science and Engineering Center (MRSEC; NSF DMR 2308708) at UC Santa Barbara. F.O. was partially funded under the National Recovery and Resilience Plan (NRRP), Mission 4 Component 2 Investment 1.4–Call for tender no. 3138 of 16 December 2021, rectified by decree no. 3175 of 18 December 2021 of the Italian Ministry of University and Research funded by the European Union–NextGenerationEU (Project code CNO0000033, Concession decree no. 1034 of 17 June 2022 adopted by the Italian Ministry of University and Research, CUP D43C22001280006, Project title “National Biodiversity Future Center–NBFC”). Views and opinions expressed are, however, those of the author(s) only and do not necessarily reflect those of the European Union or the European Commission, nor can they be held responsible for them. Additional funding sources for individual projects was provided by the following sources (in alphabetical order): Administration for Native Americans, Arctic Goose Joint Venture, Ayers Wild Cat Conservation Fund, The Bailey Wildlife Foundation, Berkeley Fellowship, Bowhunters of Wyoming, Buffalo Bill Center of the West, Bureau of Land Management, California Department of Fish and Wildlife, California Department of Water Resources, California Rice Commission, Coastal Raptors, Cooperative Agreement Number U01CK000509-01 between the Centers for Disease Control and Prevention and Northeast Regional Center for Excellence in Vector Borne Diseases, Deutsche Forschungsgemeinschaft (DFG, German Research Foundation) under Germany’s Excellence Strategy (EXC 2117–422037984), Donors to Audubon Canyon Ranch, Ducks Unlimited, European Union’s Horizon 2020 Research and Innovation program under the Marie

Sklodowska-Curie grant agreement (no. 798091), Federal Pittman–Robertson Wildlife Restoration Act, Fonds de recherche du Qu  bec Nature et Technologies (271544), George B. Storer Foundation, Georgia Department of Natural Resources–Wildlife Resources Division, Graduate Group in Ecology Student Endowment, Grand Teton National Park Foundation, Greater Yellowstone Coalition, Hawk Mountain Sanctuary, Hellman Family Faculty Fund, Henry A. Jastro Research Fellowship, James Ridgeway Endowment–University of Washington, John and Adrienne Mars, Kenneth Molson Foundation, Knobloch Family Foundation, Louisiana Department of Wildlife and Fisheries, Louisiana State University Agricultural Center, Maine Outdoor Heritage Fund, Manitoba Fish and Wildlife Enhancement Fund, Mildred E. Mathias Graduate Student Research Grant, Minnesota Department of Natural Resources, Minnesota Environmental and Natural Resources Trust Fund as recommended by the Legislative–Citizen Commission on Minnesota Resources (LCCMR), Muley Fanatic Foundation, National Geographic Society (including NGS-61630R-19 & WW-100C-17), Natural Sciences and Engineering Council of Canada (RGPIN- 2019-05185), Natural Sciences and Engineering Research Council of Canada, National Science Foundation’s Coupled Natural Human Systems 2/Dynamics of Integrated Socio-Environmental Systems (CNH2/DISES) program (Award no. 1924061), National Science Foundation Division of Environmental Biology (1652420), National Wild Turkey Federation, Nebraska Agricultural Experiment Station with funding from the Hatch Act through the United States Department of Agriculture–National Institute of Food and Agriculture, Nebraska Game and Parks Commission through the Federal Aid in Wildlife Restoration Act (W-127-R-1), Nevada Department of Wildlife, New York City Department of Parks and Recreation, Noble Research Institute, LLC, Pennsylvania Department of Military and Veterans Affairs, Bureau of Environmental Management, Private donation, Rocky Mountain Elk Foundation, Samuel Freeman Charitable Foundation, Sarah K. de Coizart Article TENTH Perpetual Charitable Trust, Schmidt Mentorship Award, Sentinelle Nord program from the Canada First Research Excellence Fund and the Network of Centre of Excellence of Canada (ArcticNET; PP12), School of Natural Resources at the University of Nebraska–Lincoln, South Carolina Department of Natural Resources, Teton Conservation District, Teton County Wyoming Government, Texas Comptroller of Public Accounts: Interagency Cooperation Contract 6670CS and Extension, Texas Parks and Wildlife Department, US Department of Agriculture, US Department of Agriculture–Forest Service, US Department of Agriculture–National Wildlife Research Center, US Fish and Wildlife Service, US Fish and Wildlife Service Webless Migratory Game Bird Program, US Geological Survey, US Geological Survey (Science Support Program) through Research Work Order no. 101 at the US Geological Survey Minnesota Cooperative Fish and Wildlife Research Unit, Utah Department of Natural Resources, Utah Division of Wildlife Resources, Utah State University, Veterinary Initiative for Endangered Wildlife, Virginia Department of Game and Inland Fisheries (PSFNP022), Virginia Tech University, Wallace Research Foundation, White Rock and Surrey Naturalists, Wildlife Restoration aid grant (WR87-R44/5), The Wild Sheep Foundation, Wyoming Animal Damage Management Board, Wyoming Department of Transportation, Wyoming Game and Fish Department, Wyoming Governor’s Big Game License Coalition, The Wyoming Wildlife Livestock Disease Research Partnership, Wyoming Wildlife & Natural Resource Trust, Wyoming Wild Sheep Foundation, Yellowstone Forever. **Author contributions:** R.Y.O., S.W.Y., D.E.S., B.R.J., C.R., and W.J. contributed to conceptualization. D.E.S. and J.C. led data curation with contributions from B.R.J., S.G., B.S.C., S.C.D., F.O., and J.S. R.Y.O. and S.W.Y. designed methodology and performed formal analysis. T.A., K.B., N.W.B., D.B., K.B., J.G.B., G.B.R., J.L.B., J.F.B., J.B., D.E.B., D.B., N.B., A.J.B., B.S.C., M.L.C., M.J.C., B.A.C., A.C., D.D. V.D., C.R.D., R.C.D., D.D., L.M.E., J.E., B.A.E., W.M.F., D.H., M.H.M., J.E.H., M.R.J., S.J., M.J.K., R.K., M.J.K., B.M.K., M.L., S.L., T.N.L., J.L., P.L., M.C.L., D.L., L.A.M., J.M.M., D.M., F.M., K.L.M., L.N., A.C.O., C.O., J.C.P., T.R.P., L.P., K.A.S.F., M.S., A.S., S.S., R.A.S., J.S., D.R.S., J.S., R.D.S., N.J.S., J.F.T., P.T., M.V., E.V.W., D.V.M., T.V., B.L.W., N.W., S.L.W., C.W., C.C.W., D.W.W., and J.Y. contributed to the investigation by providing datasets. R.Y.O. led the writing of the original manuscript draft with substantial contributions from S.W.Y. and C.R. R.Y.O., S.W.Y., D.E.S., B.R.J., J.C., S.G., R.P. T.A., K.B., N.W.B., D.B., K.B., J.G.B., G.B.R., J.L.B., J.F.B., J.B., D.E.B., D.B., N.B., A.J.B., B.S.C., M.L.C., M.J.C., M.J.C., B.A.C., A.C., S.C.D., and D.D. V.D., C.R.D., R.C.D., D.D., L.M.E., J.E., B.A.E., W.M.F., D.H., M.H.M., J.E.H., M.R.J., S.J., M.J.K., R.K., M.J.K., B.M.K., M.L., S.L., T.N.L., J.L., P.L., M.C.L., D.L., L.A.M., J.M.M., D.M., F.M., T.M., K.L.M., T.M., L.N., A.C.O., F.O., C.O., J.C.P., T.R.P., L.P., K.A.S.F., M.S., A.S., S.S., R.A.S., J.S., D.R.S., J.S., R.D.S., N.J.S., J.F.T., P.T., M.V., E.V.W., D.V.M., T.V., B.L.W., N.W., S.L.W., C.W., C.C.W., D.W.W., J.Y., C.R., and W.J. contributed to reviewing and editing of the manuscript draft. **Competing interests:** C.R. holds advisory roles relating to the topic of movement ecology, with the National Geographic Society, the International Bio-Logging Society, UNEP’s Convention on the Conservation of Migratory Species of Wild Animals (CMS), WILDLABS, MoveBON, and Save the Elephants’ Continental Tracking Initiative. R.Y.O., S.W.Y., and C.R. hold advisory roles with or participate in MoveBON. The authors declare no other competing interests. **Data, code, and materials availability:** All data, including environmental annotations, area-normalized human mobility data, and estimated area and niche size as well as code to reproduce results are available on the Open Science Framework (50). Animal tracking datasets are available on Movebank Repository, DataDryad, or USGS Science Base through individual data publications (51–83). A subset of animal tracking datasets are not publicly available because of conservation concern but will be made available for research purposes on request (see table S1 for contact information for each dataset). Human modification data are publicly available on Figshare (84). **License information:** Copyright    2026 the authors, some rights reserved; exclusive licensee American Association for the Advancement of Science. No claim to original US government works. <https://www.sciencemag.org/about/science-licenses-journal-article-reuse>

SUPPLEMENTARY MATERIALS

science.org/doi/10.1126/science.adq3396
Materials and Methods; Supplementary Text; Figs. S1 to S80; Tables S1 to S3; References (85–122); Reproducibility Checklist

Submitted 21 May 2024; resubmitted 1 July 2025; accepted 11 March 2026

[10.1126/science.adq3396](https://doi.org/10.1126/science.adq3396)

Interacting effects of human presence and landscape modification on birds and mammals

Ruth Y. Oliver, Scott W. Yanco, Diego Ellis-Soto, Brett R. Jesmer, Juliet Cohen, Song Gao, Robert Patchett, Tal Avgar, Keith Bildstein, Nicholas W. Bakner, David Barber, Kristin Barker, Joseph G. Barnes, Guillaume Bastille-Rousseau, Jerrold L. Belant, John F. Benson, Joël Bêty, Dean E. Beyer, Jr., David Bird, Nathaniel Bowersock, Andy J. Boyce, Ben S. Carlson, Michael L. Casazza, Michael J. Chamberlain, Michael J. Cherry, Bret A. Collier, Alyson Courtemanch, Sarah C. Davidson, Darren DeBloois, Vickie DeNicola, Christopher R. DeSorbo, Robert C. Dowler, Daniel Dupont, L. Mark Elbroch, John Elliott, Betsy A. Evans, W. Mark Ford, David Hancock, Molly Hardesty-Moore, Jason E. Hawley, Mackenzie R. Jeffress, Scott Jennings, Matthew J. Kauffman, Roland Kays, Marcella J. Kelly, Bryan M. Kluever, Myles Lamont, Scott LaPoint, Tayler N. LaSharr, Josee Lefebvre, Pierre Legagneux, Matthias-Claudio Loretto, David Lumpkin, Lindsay A. Martinez, John M. Marzluff, Douglas McCauley, Fiona McDuie, Tony W. Mong, Kevin L. Monteith, Thomas Mueller, Levi Newediuk, Anna C. Ortega, Federico Ossi, Cory Overton, J. Clint Perkins, Tyler R. Petroelje, Laura Prugh, Kimberly A. Sager-Fradkin, Michael Seer, Avery L. Shawler, Shannon Skalos, Rachel A. Smiley, Julia Sommerfeld, Daniel R. Stahler, John A. Stephenson, Richard D. Stevens, Nathan J. Svoboda, Jean-Francois Therrien, Phillippe J. Thomas, Meredith VanAcker, Eric Vander Wal, Dan E. Varland, Tana L. Verzuh, Brittany L. Wagler, Nils Warnock, Stephen L. Webb, Christopher K. Williams, Christopher C. Wilmers, David W. Wolfson, Julie K. Young, Christian Rutz, and Walter Jetz

Science **392** (6800), . DOI: 10.1126/science.adq3396

Editor's summary

Agriculture, settlements, roads, and other infrastructure profoundly shape the habitats that are available to nonhuman animals. In addition to changing their behavior to avoid or exploit these new hazards and resources, many species may also directly avoid humans. Oliver *et al.* examined how bird and mammal habitat use depends on both human land use and direct human presence using mobile device data to determine where people are (see the Perspective by Beaudrot). Across 37 species with geolocation data in the United States, most species responded directly to human movements, with most of these showing interactive effects between human land use and human presence. Considering only human effects on land cover will thus often provide inaccurate assessments of species' habitat use.

—Bianca Lopez

View the article online

<https://www.science.org/doi/10.1126/science.adq3396>

Permissions

<https://www.science.org/help/reprints-and-permissions>

Use of this article is subject to the [Terms of service](#)

Science (ISSN 1095-9203) is published by the American Association for the Advancement of Science. 1200 New York Avenue NW, Washington, DC 20005. The title *Science* is a registered trademark of AAAS.

Copyright © 2026 The Authors, some rights reserved; exclusive licensee American Association for the Advancement of Science. No claim to original U.S. Government Works